ABORIGINAL OVERKILL

The Role of Native Americans in Structuring Western Ecosystems

Charles E. Kay
Utalı State University

٠. .

em concept of wilderness as areas without human influence is a myth. ulation" approach by today's land managers will not duplicate the econities. Because ecosystems with native peoples are entirely different limiting ungulate numbers and purposefully modifying the vegetation servation strategy. Native Americans acted in ways that maximized their cans had no effective conservation practices, and the manner in which ciency of Native American predation, including cooperative hunting, use North America was not a "wilderness" waiting to be discovered, instead logical conditions under which those ecosystems developed. The modthan those lacking aboriginal populations, a "hands-off" or "natural regwith fire, Native Americans structured entire plant and animal commuhumans, conservation is seldom an evolutionarily stable strategy. By individual fitness regardless of the impact on the environment. For they harvested ungulates was the exact opposite of any predicted consynergism between aboriginal and carnivore predation. Native Amerikills; (5) impact of European diseases on aboriginal populations; and (6) reservoirs; (4) species ratios, and sex and age of aboriginal ungulate (2) optimal-foraging studies; (3) tribal territory boundary zones as prey of dogs, food storage, use of nonungulate foods, and hunting methods; Intermountain West. This hypothesis is based on analyses of (1) the effimajor factor limiting the numbers and distribution of ungulates in the Prior to European influence, predation by Native Americans was the

Received February 10, 1994; accepted April 8, 1994.

Address all correspondence to Charles E. Kay, Institute of Political Economy, Utah State University, Legan, UT 84312.

Copyright © 1994 by Walter de Gruyter, Inc. New York Human Nature, Vol. 5, No. 4, pp. 359-398.

1045-6767/94/\$1.00+.10

it was home to tens of millions of aboriginal peoples before European-introduced diseases decimated their numbers.

KEY WORDS: Aboriginal overkill; Aboriginal subsistence; Aboriginal conservation; Ungulate ecology; Yellowstone National Park (Wyoming); Precolumbian ecosystems; Predation; Predatorprey ecology; Native Americans; Ecosystem management.

LESSONS FROM YELLOWSTONE

support its paradigm (Boyce 1991; Coughenour and Singer 1991). off," "let-nature-take-its-course" management (Despain et al. 1986; Housopposition, though, the Park Service abandoned its control program in destroying Yellowstone that from 1949 to 1968 rangers shot more than ly the Park Service has cited Caughley's (1976) plant-herbivore model to logistic-growth equation and a presumed balance of nature, but recentton 1982). Agency biologists initially based "natural regulation" on the 13,500 elk to reduce the northern herd. Faced with mounting political winter range. In fact, the Park Service was so convinced elk were severely overgrazing the park, and in particular Yellowstone's northern however, concerns grew that an unnaturally large elk population was wolves (Canis lupus) and mountain lions (Felis concolor). By the late 1920s, 1968 and by the early 1970s switched to "natural regulation" or "handsworld's first national park in 1872, government officials did not think Yellowstone National Park. When Yellowstone was established as the (Bison bison), and other ungulates, and they killed predators such as that there were enough game animals, so they fed wintering elk, bison This research began as a study of elk (Cervus elaphus) and vegetation in

Under "natural regulation," predation is assumed to be an assisting but nonessential adjunct to the regulation of ungulates through density-dependent homeostatic mechanisms. The population and distribution of elk and other wild ungulates are limited by food, and according to the Park Service, it is natural for thousands of animals to starve to death. If wolves or other predators were present, they would only kill animals slated by nature to die of other causes and would not limit or lower ungulate numbers. In the current debate over reintroducing wolves to Yellowstone, the Park Service has never said wolves are needed to control the elk herd, and in fact, the agency adamantly denies that wolves will have any significant impact on Yellowstone's game populations.²

The Park Service also denies that Yellowstone was ever or is now overgrazed. Today, the agency contends that large numbers of elk

(12–15,000+) have wintered on the park's northern range for the past 8–10,000 years and that those animals have been in equilibrium with Yellowstone's plant communities. According to the Park Service, any recent (1872–1990) vegetation changes are due primarily to suppression of lightning fires, normal plant succession, or climatic change, not ungulate grazing. Under "natural regulation," elk browsing and high-lining (stripping of lower branches for food) of Yellowstone's vegetation are natural and represent the pristine condition of the park. The agency also steadfastly maintains that Yellowstone's elk have not competitively excluded sympatric herbivores, such as smaller ungulates or beaver (Castor canadensis).

inhabited Yellowstone during the 1800s, a conclusion supported by the food and dam building materials (Kay 1990, 1994; Chadde and Kay extinct because elk have eliminated willows and aspen beaver need for they were throughout the West ca. 1800, but they are now ecologically were once exceedingly common on Yellowstone's northern range, as ed ungulate browsing, not other factors (Kay 1985, 1987, 1990, 1993a; and aspen (Populus tremuloides) have declined by 95% owing to repeatthe "natural regulation" paradigm. Since 1872, tall willows (Salix spp.) park's early explorers. Park was established, another indication that few elk or other ungulates proof exclosures today look like those plants did when Yellowstone (Kay and Wagner 1991). In fact, willows and aspen inside ungulatebranches down to the ground, and aspen and willows were unbrowsed photographs, though, show that in 1871 Yellowstone's conifers had least palatable species, have been high-lined by starving elk. Historical 1991). Today, even Engelmann spruce (Picea engelmannii), one of the Chadde and Kay 1988, 1991; Kay and Chadde 1992; Patten 1993). Beaver Recent studies, however, have failed to find any evidence supporting

Between 1835 and 1876, 20 different parties spent a total of 765 days traveling through the Yellowstone ecosystem on foot or horseback, yet they reported seeing bison only three times—now there are 3,000 to 4,000 bison in the park. Elk were seen by explorers on average only once every 18 days—today there are nearly 60,000 elk in the ecosystem. The fact that many parties broke into small groups and spread out to hunt makes these observation rates all the more meager (Kay 1990). Moreover, while these explorers were in Yellowstone, their first-person journals contain 45 references to a lack of game or a shortage of food. In addition, none of the early parties reported seeing or killing even a single wolf, another indication that ungulates were scarce (Kay 1992a).

Archaeological data suggest that elk were also rare prehistorically. As outlined above, the Park Service's "natural regulation" paradigm assumes that large numbers of elk wintered on Yellowstone's northern

absent from archaeological sites in the Yellowstone area (Table 1) and square kilometer. Assuming that Native Americans killed ungulates in 52,000 unearthed ungulate bones (Wright 1984; Kay 1990, 1992b). throughout the Intermountain West, representing only 3% of more than logical sites should be dominated by elk. Elk, however, are rare or even nities had the same composition as today, then Yellowstone's archaeoproportion to their abundance and that precolumbian ungulate commuon the park's northern range at winter densities of 20 to 40 elk per the Yellowstone ecosystem today are elk, with 15,000 to 25,000 animals dominated the ungulate community. Eighty percent of the ungulates in dance of ungulate species has not changed over time; that is, elk always range for the past several thousand years and that the relative abun-

simply because they were a plains animal. cessing sites could exist but have never been excavated by archaeolocould not kill elk, or (2) they may have chosen not to kill elk. (3) Or pergists. (6) And finally, perhaps elk were not present in the mountains preservation or other taphonomic factors. (5) Then too, special elk proproblem. (4) Or conceivably this could have been caused by differentia haps elk bones were not brought to sites, inferring a transportation ty of elk in the archaeological record. (1) Perhaps Native Americans At least six possible factors can be advanced to account for the scarci-

of animals commonly subject to differential transportation (e.g., O'Congatherers have shown that only the largest bull elk fall within the size from bison kills but "ditched" elk bones. (c) Studies of modern hunterimprobable that aboriginal hunters would have brought back bones archaeological sites. Since bison are nearly twice as large as elk, it is late wintering areas so "ditching" bones would probably not have been an overriding consideration. (b) Bison bones outnumber elk bones in campsites, would often leave behind lower quality bones in favor of Many archaeological sites are found in close proximity to known unguhunters, faced with carrying portions of a large ungulate back to distant tion. Binford (1978, 1981) and others have suggested that aboriginal discussed below.) Third, the relative scarcity of elk bones in archaeoand human preferences for meat dispel any notion that native people in species for the past 10,000 or so years. Second, optimal-foraging theory mountain archaeological bone deposits for at least three reasons. (a) transporting meat. It is unlikely, however, that this factor skewed interlogical sites does not appear to be the result of differential transportathe Intermountain West chose not to kill elk. (Both of these findings are Their technology has been more than sufficient to kill all ungulate First, no evidence exists that Native Americans could not kill elk

Fourth, the suggestion that differential preservation may somehow

Table 1. The Relative Abundance of Present Ungulate Populations Compared with the Relative Abundance of Ungulate Faunal Remains Unearthed from Archaeological Sites in the Same Areas.

		- fuels	pecies F	pecies Percentage of Total	Total	
Area, Data Set	Elk	Bison	Advalle deca	Antelope	Bighorn	Moose
Greater Yellowstone Ecosystem						
Current ungulate population	79	حناس	9	1	42	حلم
Archaeological sites (MNI)	Οì	15	29	÷	16	0
Sunlight Basin*						
Current ungulate population	72	0	10	0	14	42
Dead Indian Creek site (NISP)	_	حلم	67	2	26	0
North Fork of Shoshone*						
Current ungulate population	65	0	19	0	<u>_</u>	2
Mummy Cave site (MNI)	1		14	0	\$	0
Archaeological sites (MNI=3,345)	4	=	37	17	8	0.030
Archaeological sites (NISP=52,624)	ω	တ	56	9	23	0.002

cast of Yellowstone Park. See Prison and Walker (1984) and Hatris (1978).

Adapted from Kay (1990). MNI = minimum number of individuals, NISP = number of identi-

is also no evidence that aboriginal peoples treated elk bones differently mountain sites. Based on butchering marks and breakage patterns, there exact opposite of the species-abundance patterns observed at interothers have shown that, in general, large dense bones preserve better ed by available evidence. Taphonomic studies by Binford (1981) and from those of other ungulates. (Odocoileus hemionus) and bighorn sheep (Ovis canadensis) bones, the preservation should favor elk and work to the detriment of mule deer than small light bones. Based on these considerations, differential explain the scarcity of elk in archaeological contexts is also not support-

a bias against elk processing sites doubtful even if the latter existed. to base camps and permanent villages—that have been excavated make many types of archaeological sites-from temporary camps to kill sites gests that this cannot account for the observed pattern. Furthermore, the dismissed, the large number of documented habitation sites strongly sugnever been excavated or found by archaeologists cannot be summarily Fifth, although the idea that special elk-processing sites exist but have

was primarily a plains animal, but this supposition is not supported by ecological data. Biological studies on digestive efficiency, diet breadth, did not inhabit the mountains in prehistoric times because the species both Frison (1991) and Wright (1984) conclude that large numbers of elk Sixth, based on their archaeological experience in western Wyoming,

and energetics have all shown that elk are superior competitors to bighorn sheep and mule deer on intermountain winter ranges. Elk will simply outcompete, and outnumber, the smaller ungulates. If elk thrive in the Yellowstone ecosystem and other western mountains today, why were they rare in prehistoric times?

Thus, vegetation data, early photographs, historical first-person journals, and archaeological data all suggest that large numbers of resource-limited ungulates did not inhabit the Yellowstone ecosystem until the onset of European influence. But what is different today, and what could account for this change? Unlike much of North America, the Yellowstone ecosystem has remained relatively intact, except that wolves and Native Americans are no longer present. While early records suggest that wolves were also rare in Yellowstone, first-person journals written between 1835 and 1876 contain 53 references to Native Americans (Kay 1990, 1992a).¹ This led me to consider the role native peoples played in prehistoric ungulate ecology. That research forced me to conclude that prior to European influence, predation by Native Americans limited the numbers and distribution of ungulates in the Yellowstone ecosystem and throughout the Intermountain West.*

Although the demonstrated lack of elk in archaeological sites may at first appear to negate my aboriginal overkill hypothesis, in fact, the opposite is true. Optimal-foraging theory (see below) predicts that high-ranked items, like elk or other ungulates, are more susceptible to over-exploitation than low-ranked items, such as vegetal foods, small mammals, or fish. According to optimal-foraging models, high-ranked items will seldom appear in the diet if they are being overexploited. So, ungulate species unearthed with the lowest frequency in archaeological sites, such as moose (Alces alces, represented by only 1 of more than 52,000 bones) and elk, were probably subjected to extreme overexploitation. Moreover, the small proportion of large mammals in intermountain aboriginal diets, both historically and prehistorically, as well as the highly fragmented nature of archaeologically recovered bone suggest that all species of ungulates were relatively rare for the past 10,000 years.

EVIDENCE THAT PRECOLUMBIAN UNGULATE POPULATIONS WERE LOW

Besides archaeological data, four additional lines of evidence suggest that precolumbian ungulate populations were low and that they were not limited by food: historical journals, aboriginal use of berries such as serviceberry (Amelanchier alnifolia), types of wood used for arrow shafts, and tribal buffer zones. First, in addition to the Yellowstone journals dis-

cussed above, I have also conducted continuous-time analyses on approximately 100 first-person accounts of the Intermountain West written between 1790 and 1850. Except as noted below, those journals all indicate that game was seldom encountered in the western mountains. For instance, between 1792 and 1871, 21 different parties spent 369 days in the southern Canadian Rockies but reported seeing elk only 12 times (Kay et al. 1994).

stream bordered with black cherry trees [chokecherries], many of the signs" (Langford 1972:13). Since shrubs have to be at least 2 m tall before smaller ones broken down by bears, of which animal we found many dition of 1870 reported that near Yellowstone Park "we crossed a small three bushels drying in the sun" (Haines 1965:16). The Washburn Expeengaged in gathering and drying choke-cherries ... they had two or cans who were gathering and drying large quantities of chokecherries at plants in 1870 not only produced abundant berries but were also very branches are commonly broken down by feeding bears, chokecherry Park. "Here we found a wickiup inhabited by two old squaws who were the mouth of Tom Miner Creek a few kilometers north of Yellowstone 1869, the Cook-Folsom-Peterson Expedition encountered Native Amerias serviceberries and chokecherries (Prunus virginiana). In September Native Americans commonly consumed large quantities of berries, such Chamberlin 1911; Lowie 1924) and archaeological studies reveal that A second line of evidence is berry production. Ethnohistoric (e.g.

Conditions today are very different. Serviceberry and chokecherry plants in Yellowstone are now less than 50 cm tall and they produce virtually no berries because the plants are repeatedly browsed by elk and other ungulates (Table 2). Resource-limited ungulate populations and large quantities of berries are mutually exclusive on western ranges. Even moderate numbers of ungulates curtail berry production because these plants provide highly preferred forage, especially in winter. Ungulate-induced berry reduction is even reflected in grizzly bear (Ursus arctos) food habits. Whereas grizzlies in Canada and Alaska commonly consume large quantities of berries, bears in the Yellowstone ecosystem do not. From 1977 to 1992 more than 10,000 grizzly bear scats were collected and analyzed in Yellowstone, yet chokecherries were only reported in one scat, serviceberries in two, and buffaloberries (Shepherdia canadensis) in 51 (Kay 1993b). The fact that Native Americans in the West consumed large quantities of berries both historically and prehistorically means that ungulate numbers were low and those populations were not limited by food.

A third line of evidence is plant growth form. Ethnographic accounts and archaeological finds indicate that Native Americans preferred to

lowstone Ecosystem. The number of berries produced by plants protected from browsing inside ungulate-proof exclosures compared with the number of berries produced by the same species outside the exclosures. The Effect of Ungulate Browsing on Berry Production in the Yel

	Number of Berries per 100 Plants	s per 100 Plants	
Exclosure, Species	Inside	Outside	טי
Camp Creek			
Serviceberry	133,307	7	<.001
Lamar-west			
Serviceberry	111,047	0	<.001
Chokecherry Uhl Hill	212,178	0	<.001
Serviceberry	10,468	0	<.001
Chokecherry	6,508	0	<.001
Mammoth			
Buffaloberry	119,146	250	<.001
Total	592,654	257	<.001

Adapted from Kay (1993b)

(Kay 1993b). This is also true on other ranges throughout the West. A fourth line of evidence is aboriginal buffer zones. Mech (1977, 1994) in Yellowstone today, except where plants are protected from ungulates severely hedge these plants because serviceberry is so highly preferred. straight enough for arrows. Even moderate numbers of ungulates It is impossible to make arrows from any of the serviceberry that exist able serviceberry makes it difficult to find branches that are long and Sinopoli 1991). Now, however, repeated browsing of the highly palatdurable and those plants grew straight and tall, especially under aboriginal management such as pruning or burning (e.g., Anderson 1991) make arrows from serviceberry branches because the wood was very

entitled "The Virginia Deer and Intertribal Buffer Zones in the Upper buffer zone phenomena in the literature, a paper by Hickerson (1965) ginianus) to survive in those areas and to live longer (Hoskinson and quently than the central part of their ranges in order to avoid encoun-Mississippi Valley." Hickerson (1965:45) noted that Mech 1976). Mech (1977) could find only one other instance of this the territorial edges, which permitted more white-tailed deer (O. virters with neighboring wolves. This reduced predation pressure along reported that wolf packs used the edges of their territories less fre-

ing hunters from occupying the best game region intensively enough to Warfare between members of the two tribes had the effect of prevent-

Aboriginal Overkill

a protective zone for the deer, was destroyed and famine ensued was maintained between certain Chippewa and Sioux, the buffer, in effect deplete the (deer) supply. . . . In the one instance in which a lengthy truce

zones created by Native American hunting. Lewis and Clark (1893) and no game. Raynolds (1868:38) noted that grass, whereas to the west he again encountered an abundance of grass Powder River, though, he reported an abundance of game and little an abundance of grass but no game east of the Powder River. Along the the country lying between nations at war." In 1859, General Raynolds observe that the greatest quantities of wild animals are usually found in 1197), for instance, noted that "With regard to game in general, we (1868), who led an expedition across the Dakotas and Montana, found My research, however, has uncovered frequent references to bufter

species of neutral ground between the Sioux and the Crows and other ren region [Powder River] is explained by the fact that this valley is a son the buffalo remain here undisturbed and indeed would seem to make would thereby give notice to their enemies of their presence. For this reaground visited only by war parties, who never disturb the game, as they bands nearer the mountains, or, more correctly speaking, the common war the valley a place of refuge. The presence of these animals [bison] in such large numbers in this bar-

other carnivores (see below) was not the primary factor limiting preof aboriginal buffer zones also indicates that predation by wolves and suggest that aboriginal hunting limited ungulate numbers. The presence only were ungulate populations generally low, but these data strongly modern hunter-gatherers (e.g., Smith and Winterhalder 1992). This pataround camps and villages has frequently been reported in studies of or to drive out game animals (e.g., Post 1938:11), and resource depletion columbian ungulate populations abundance of game in a few locations and a lack of game elsewhere. Not boundary zones also explain how early explorers could encounter an with no effective conservation practices (see below). Tribal territory tern would be expected if people pursued an optimal-foraging strategy Historical sources indicate that aboriginal hunting tended to extirpate

HOW ABORIGINAL HUNTERS LIMITED UNGULATE POPULATIONS

means, including cooperative hunting; use of drives, traps, and corrals; Aboriginal hunters could limit the numbers of ungulates by several

369

use of dogs; weapons that kill at a distance; long-distance pursuit; use of snowshoes; food storage; and the use of fire.

(e.g., Hawkes 1987; Hawkes and O'Connell 1981; Hawkes et al. 1985). themselves, and they certainly do not live a hand-to-mouth existence day hunter-gatherers still spend relatively little time provisioning mates for !Kung work effort were too low, the !Kung and other presentety." Although more recent studies have shown that Lee's original esti-(1972) went so far as to call hunter-gatherers "the original affluent sociperson living in today's "most advanced" western civilizations. Sahlins gested that "primitive" people had more leisure time than the average living in one of the most inhospitable environments on Earth. Lee suged that the !Kung spent relatively little time in the quest for food despite "Man the Hunter" conference (Lee and DeVore 1968). Lee demonstratthe publication of Lee's (1968) research on the !Kung and the subsequent abandoned this stereotype of "primitive" people three decades ago with Americans as important ecological factors. Anthropologists, however, for food underlies most biologists' out-of-hand dismissal of Native lived a brutish existence and spent every waking moment in the quest Although not often explicitly stated, the idea that prehistoric humans

Numerous ethnohistoric accounts indicate that Native Americans commonly ran down ungulates (e.g., Anell 1969). Carrier (1984) has even suggested that humans evolved as long-distance endurance predators. Wolves and mountain lions, on the other hand, seldom chase ungulates more than 400 to 800 m. Where there is a differential accumulation of snow in western mountains, Native Americans on snowshoes could simply run ungulates uphill into deeper and deeper snow where they were able to kill the floundering animals, including elk, often with no more than handaxes or clubs (e.g., Lewis and Clark 1893:623; Smith 1974:54–55).

In addition, Native Americans used dogs to hunt all species of ungulates. Because of its effectiveness, that practice has been outlawed in all western states since the inception of modern game management. In the Kalahari, one !Kung with trained hunting dogs brought in 75% of the total meat obtained by one camp, while six hunters who lacked dogs accounted for the remaining 25% (Washburn and Lancaster 1968: 294–295). Not only did Native Americans hunt in a truly cooperative manner, they also employed various drives (including fire drives), traps, and corrals to take all species of ungulates (e.g., Anell 1969). Whether they used spears, atlatls, or bows and arrows, aboriginal peoples in North America have always killed at a distance, thereby reducing the risk of physical injury that carnivorous predators face each time they attempt a kill. Powerful sinew-backed horn bows, common to many intermountain tribes, could drive arrows completely through even the largest ungulates (Townsend 1978).

Humans also derive an advantage from the division of labor in hunter-gatherer societies where males hunt and females gather mainly vegetal foods. Male wolves do not kill with tools, butcher, transport, and share with females who have been gathering other foods that are in turned shared with males. Unlike carnivorous predators, humans usually extend considerable care to sick or injured group members. The success of the hunting and gathering way of life lay in adaptability that permitted a single species to occupy most of the Earth with a minimum of biological adaptation to local conditions (Washburn and Lancaster 1968).

Aboriginal peoples' ability to kill ungulates depended not only on their behavior but also on the prey's. For instance, moose that stand and hold their ground when tested by wolves have a higher probability of survival than individuals that attempt to flee (Peterson 1977). This behavioral strategy, evolved through eons of coevolution, may be adaptive when moose encounter wolves, but the same strategy is fatal when moose are hunted by Native Americans who kill at a distance. Aboriginal peoples used dogs to bay moose in order to take full advantage of this situation. This combination of factors made killing moose extremely easy, despite those animals' large size.

Clearly, native peoples possessed the technical skills and physical means to kill any and all ungulate species, often at will. Native Americans armed with no more than spears, for instance, were able to kill even grizzly bears with little difficulty or risk to human life (Birkedal 1993).

THEORETICAL CONSIDERATIONS SUPPORTING ABORIGINAL LIMITATION OF UNGULATE POPULATIONS

There are also several theoretical considerations that support the hypothesis for aboriginal limitation of ungulate populations, including prey switching and the use of nonungulate foods, optimal-foraging theory, age structure of the kills, sex of the kills, synergism between aboriginal and carnivore predation, and the lack of effective aboriginal conservation practices.

Prey Switching

According to predation theory, a single predatory species cannot take a single prey species to extinction because the energy required to locate the last remaining prey is more than the predator would receive from securing those prey items. Bergerud (1983) demonstrated, however, that prey switching by a single predatory species in a relatively simple

ecosystem drove one prey species to near extinction, damped the cyclic fluctuations of a second prey species, and limited the population of a third species, an ungulate. The predator maintained a higher population level by feeding on three prey species than by subsisting on only a single prey. In addition, the predator exerted more predatory pressure on all prey species than it could have exerted if only one had been available.

Prey switching by wolves between age classes of prey, between ungulate species, and even to smaller mammals, like beaver, has been widely reported. Messier (1985) and Messier and Crete (1985) suggest that prey switching to nonungulate foods enabled wolves in Ontario to subsist despite low moose densities. In fact, the wolf populations that they studied would not have been able to survive without nonungulate alternative foods. Voigt et al. (1976) also reported that alternative prey permitted wolves to maintain relatively stable, high-density populations, while Fuller and Keith (1980) found that garbage dumps enabled wolves to achieve higher densities than was otherwise possible.

Unlike purely carnivorous predators, humans can switch from large game to small mammals, fish, insects, or vegetal foods, and ultimately to agriculture. Contrary to the notion that Native American diets were primarily meat (McCabe and McCabe 1984:28), anthropologists have long noted that aboriginal peoples should more appropriately be called gatherer-hunters. Except for Arctic Eskimos and perhaps Plains tribes after the introduction of the horse, vegetal foods and fish comprised 80% to 90% of historical and prehistoric diets, especially in the Intermountain West. Native Americans preferred meat when it was available, however (Webster 1983:44); vegetal foods ranked a poor second despite their high nutritional value (e.g., Gould 1982:77). Moreover, all cultures accorded hunting more prestige and status than they did gathering (Hawkes 1990, 1991, 1992, 1993).

By prey switching to a diet of largely fish or vegctal foods, Native American populations could have continued to grow despite the increasing scarcity of their preferred ungulate foods and diminishing returns of the hunt. In the eastern United States, Webster (1979) reports that the Huron hunted white-tailed deer despite a "considerable energetic loss." Although diminishing returns act as a homeostatic mechanism to control populations of some predators, little such control has operated in the case of humans (Cohen 1977:187). Unlike wolves, humans could severely limit or exterminate ungulates without causing a major decline in their own population because people could rely on vegetal resources or fish. Prey switching and food storage make humans perhaps the most starvation tolerant of all predators, and the more starvation tolerant a predator is, the greater the impact it can have on its preferred prey. Optimal-foraging theory supports similar conclusions.

Optimal-Foraging Theory

Optimal-foraging theory represents an attempt to develop a set of models general enough to apply to a broad range of species yet rigorous and precise enough to explain details of individual behavior. The theory assumes foraging behavior evolved by natural selection to respond to changing conditions in ways that maximize each forager's individual survival and reproductive success. Optimal-foraging theory attempts to specify a general set of decision rules for predators based on cost-benefit considerations which are in turn deducible from first principles of adaptation via natural selection (Stephens and Krebs 1986). Originally developed by biologists to study nonhuman animals, optimal-foraging models have been employed by anthropologists to examine human foraging. They have been used with success to study a number of hunter-gatherer societies, and to improve our understanding of the archaeological record (e.g., Simms 1984).

ductive success (e.g., Alvard 1993a, 1993b; Smith 1983, 1991; Smith and give ungulates added value; and (6) better hunters have higher reprothree to four times more value per calorie than plant foods; (5) hides prey-choice instead of the patch-choice model; (3) ungulates or other er, and only men hunt ungulates; (2) most food selection follows the general, concluded that (1) men usually hunt while women mostly gathungulates even more profitable to pursue and puts additional harves cies, and any reproductive advantages that superior hunters may enjoy tent, hides, and for other social considerations. These additional currenprocurement of other foods), they are also valued for their high fat conitems (i.e., ungulate hunting yields more calories per unit time than the higher-ranked, energy-based handling efficiencies than most other diet large mammals are the highest ranked resources; (4) meat and tats have pressure on them even at low population densities. have the effect of increasing ungulate handling efficiencies. This makes Winterhalder 1992, Winterhalder 1987). So not only do ungulates have Recent optimal-foraging research on modern hunter-gatherers has, in

According to optimal-foraging theory, ungulates will be taken whenever they are encountered, and a diet of low-ranked items, such as vegetal foods or fish, means that high-ranked items are rare or absent. Recall that aboriginal diets throughout the Intermountain West were high in vegetal foods, small mammals, and fish, all low-ranked items, and low in high-ranked ungulates. This implies that, historically and prehistorically, all ungulate populations were low. Optimal-foraging theory also predicts that high-ranked items are the most susceptible to overexploitation. These factors suggest that in precolumbian times Native Americans were probably overexploiting elk and other ungulates.

At winter densities of 20 to 40 elk per square kilometer, which are now common in Yellowstone Park, optimal-foraging models would predict aboriginal diets should be nearly 100% elk. But archaeologically, that did not happen. This can only mean that few elk were actually available to prehistoric hunters, and that today's ungulate population densities do not represent precolumbian conditions. This is true not only in Yellowstone, but throughout the Intermountain West.

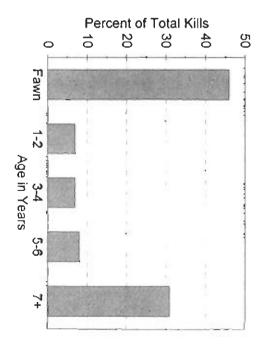
Age Structure of Aboriginal Kills

Archaeologists (e.g., Lyman 1987) have used the shape of mortality profiles to establish whether those ungulates were killed all at once or over a period of time. This catastrophic versus attritional mortality dichotomy, however, is based on the assumptions that the populations from which those samples were drawn had stable age structures and that all age classes were randomly distributed through the populations (Voorhies 1969). Unfortunately, this assumption is not valid because wild ungulate populations rarely exhibit stable age structures (Eberhardt 1987:117) and because most ungulates segregate by sex or age (e.g., Clutton-Brock et al. 1985). Although mortality profiles do not indicate how animals were killed, they do provide an estimation of predator effectiveness and relative ungulate mortality rates. A population subjected to increasing predation will be dominated by younger animals.

Ecological studies (e.g., Carbyn et al. 1993) have shown that wolves and mountain lions usually kill a disproportionate number of young-of-the-year and old ungulates (Figure 1a). Carnivores also tend to prey on sick, malnourished, or otherwise debilitated animals. Temple (1987:669) notes that "The degree to which substandard individuals of a particular prey species are taken disproportionately by a predator seems to be a direct function of how difficult it normally is for the predator to capture and kill individuals of that species." The more difficult it is for a predator to capture a particular prey, the more that predator will take substandard individuals and young. If two or more predators are preying upon the same species, the least efficient predator will tend to kill fewer prime-age animals (Okarma 1984).

While wolves and other carnivores kill primarily young-of-the-year and old animals, Native Americans killed mostly prime-age ungulates. For instance, Frison (1971) excavated a pre-horse Shoshone antelope (Antilocapra americana) processing site in Wyoming's upper Green River Basin. Of 79 antelope at the Eden-Farson site whose ages could be determined, 80% were less than 2.3 years old and none were older than 6.3 years (Nimmo 1971). During 1985, the Wyoming Game and Fish Depart-

Aboriginal Overkill



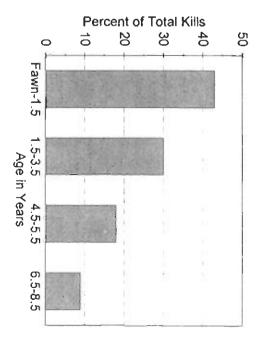


Figure 1. Age structure of ungulates killed by wolves and Native Americans. (a) Age structure of white-tailed deer killed by wolves in Minnesota (Fritts and Mech 1981). Wolves and other North American carnivores generally take a disproportionate number of very young and very old animals. (b) Age structure of mule deer unearthed from the (4200 m) Dead Indian Creek archaeological site in northwest Wyoming east of Yellowstone Park (Simpson 1984). Unlike carnivores, Native Americans killed a predominance of prime-age ungulates—an indication that Native Americans were more efficient predators. This also indicates that aboriginal peoples had a greater impact on prey population dynamics than carnivores, especially given the fact that natives killed mostly females (see text).

were under 3.5 years, and 92% were less than 5.5 years old (Figure 1b) three percent of the aged deer (n = 60) were less than 1.5 years old, 73% dominate the ungulate faunal remains from that site (Table 1). Fortyboundary of Yellowstone Park (Frison and Walker 1984). Mule deer Creek site is located in Wyoming's Sunlight Basin near the northeast heavily harvested by modern hunters. The 4,200-year-old Dead Indian Americans had a younger age structure than today's herd, which is riginally hunted animals. That is to say, the antelope killed by Native antelope included fewer fawns and more old individuals than the aboment live-trapped 333 antelope in that same area (Roper 1986). Those

egy (Hastings 1983, 1984; Michod 1979) and indicates that Native Ameranimals, however, runs contrary to any maximum sustained yield stratsuggests that, in general, Native Americans were more efficient unguusually exhibit mortality profiles dominated by prime-age animals, this icans could have had a major impact on precolumbian ungulate populate predators than wolves or other carnivores. Killing mostly prime age killed by Native Americans. lations. This is even more true when one considers the sex of ungulates Since ungulates recovered from intermountain archaeological sites

Sex of Aboriginal Kills

authors claim that Native Americans also spared females so there would and mountain lions kill a disproportionate number of males. Some will fall, often precipitously. Ecological studies have found that wolves (e.g., Short 1979). If too many adult females are killed, the population maximize sustainable harvest, few prime-age females should be killed be no diminution in the supply of game animals (Heizer 1955:4-5 ungulate populations are most sensitive to adult female mortality. To fatter and had better hides than males. Native Americans preferred to kill females because those animals were (e.g., Alvard 1993a; Teit 1928:243; Wright and Miller 1976:301). Instead ical accounts, studies of modern hunter-gatherers, or archaeological data Roberts 1932:290). This contention, however, is not supported by histor-Demographic studies (e.g., Nelson and Peek 1982) have shown that

sumption of animals fats may even have been a physiological necessity ungulates have greater stores of fat than males, especially during fall proportionate number of females because, for most of the year, female This demand for fats would have caused native hunters to kill a disother foods (e.g., Speth 1987; Speth and Spielmann 1983). The confor many Native Americans (e.g., Speth 1989, 1991; Spielmann 1989). All aboriginal people have a preference for animal fats over nearly al

> prime-age ungulates have larger fat deposits than young animals (Johns et al. 1984). and early winter (e.g., Anderson et al. 1972; Flook 1970). Moreover,

intraspecific fighting (Geist 1986; Post 1938:19). makes their hides easier to tan, and females are seldom scarred from they have better hides than males. They lack thick dermal shields, which general fullness." In addition, females were generally preferred because [moose] at a glance by their dark color, the curve of their back and their lates that have the greatest fat levels during any particular hunt. Nelson have reported that hunters tend to take the age and sex classes of ungu-(1983:165) notes that "skilled hunters can pick out the best animals Nelson (1973:98, 1983:165), Binford (1978:40), Speth (1983), and others

of ungulates harvested, Native Americans had a greater impact on prey value, such as prime-age individuals and females. For a given number icans focused their efforts on animals with the greatest reproductive offtake approaching maximum sustained yield (e.g., Hastings 1983 harvesting prime-age animals and females would ever have led to ar ber of animals. It is doubtful that the Native Americans' propensity for populations than if carnivorous predators had killed an identical numyoung-of-the-year, the old, the unfit, and males, whereas Native Amersex classes that have the lowest relative reproductive value; namely In summary, wolves and mountain lions kill primarily those age and

Carnivore Predation

ungulate numbers even lower and maintain them at that level. This overexploitation, or other causes, wolves and other predators can drive If ungulate populations have been reduced by severe weather, human satory, and together can have a major impact on ungulate numbers. (4) adult ungulates. Wolf and bear predation are additive, not compenpresent, they often prey heavily on newborn and, to a lesser degree not compensatory. (3) If grizzly or black bears (Ursus americanus) are Human and carnivore predation on ungulate populations are additive late population to predators is not enough to offset predation losses. (2) or "naturally regulated," and any compensatory response of the ungulevel set by food resources; that is, ungulates are not resource limited tions, wolves and other predators limit ungulate populations below the 1992). These studies can be summarized as follows. (1) In many situa-Carbyn et al. 1993; Gasaway et al. 1992; Messier 1991, 1994; Seip 1991, other carnivores, more often than not, limit ungulate populations (e.g., Recent research in Alaska and Canada indicates that wolves and

condition is commonly called a predator pit, and there is no field evidence that ungulates can escape from a predator pit even if hunting is banned, unless wolves and other predators are reduced. As Alaska biologists have noted, "prey [ungulate] populations can reach extremely low densities under natural conditions, contrary to the 'balance of nature' concept" (Gasaway et al. 1983:6). Throughout much of Alaska and Canada, ungulate populations are now being kept at low levels by the combined actions of carnivorous predators even in areas where they are not hunted.

Wolves and other carnivores limit ungulate numbers by reducing recruitment and increasing adult mortality, not by killing off all the game, instances of surplus killing notwithstanding. In any given year, a number of adults die from natural causes, disease, or predation. When expressed as a percentage, this is termed the adult mortality rate. In that same year, a number of calves or fawns are born, but those young also face disease, accidents, and predation, and only a few survive their first year of life to join the adult population. This is called the recruitment rate. For a stable population, recruitment, and especially female recruitment, must balance adult mortality. If recruitment is less, the population declines, and if it is greater, numbers increase (Bergerud 1990, 1992).

As indicated above, wolves and other carnivores prey most heavily on young-of-the-year, which lowers the recruitment rate of the prey populations. Predators also kill a few prime-age adults (Figure 1a). By increasing adult female mortality and at the same time lowering recruitment, carnivores can cause ungulate populations to decline. Stabilizing recruitment for caribou is about 15 female yearlings per 100 cows. Caribou herds with few predators have recruitment rates of 20 to 40 female yearlings per 100 cows, which allow those populations to increase, whereas caribou herds subject to heavy predation have recruitment rates of 10 or less. So predation causes ungulate populations to gradually decline over time—wolves do not normally wipe out game herds in a single year or two.

This slow decline is what happened in Alaska and Canada. During the 1950s and 1960s, when wolf control was widespread and effective, game herds grew and the north country became known as a hunter's paradise. Government wolf control ended by 1970, and predator populations began to expand, but it took ten or more years before significant declines were seen in game herds. In Wood Buffalo National Park, for instance, there were approximately 12,000 bison when wolf control was terminated, but today there are fewer than 3,500 and the population is still falling. Wolf predation of calves has been identified as the primary factor responsible for that decline, since the bison are not hunted (Carbyn et al. 1993).

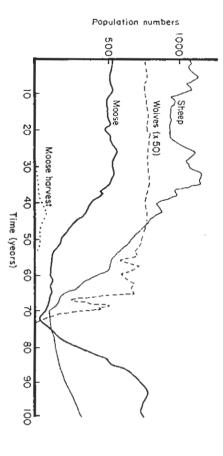
Across Canada and Alaska, moose and caribou populations not subject to heavy predation have densities ten times greater than populations where carnivore numbers are high (e.g., Gasaway et al. 1992; Messier 1994). The presence of large numbers of carnivores reduces the numbers of ungulates available for human hunters by up to 90% or more. As in the case of Wood Buffalo National Park, wolves alone can completely eliminate any "surplus" ungulates that would otherwise be available for human consumption. Moreover, if carnivore predators can limit ungulate numbers, and if they are less efficient predators than Native Americans, as I have argued, then it is easy to see how aboriginal peoples could have had a major impact on precolumbian ungulate populations.

Predator-prey models have also been developed in which carnivore and human predation act in concert on ungulate populations (Haber 1977; Walters et al. 1981). Computer simulations with these models have shown that small amounts of human predation added to wolf-bear-ungulate systems can cause the virtual collapse of both ungulate and wolf populations, even if humans are limited to killing only males (Figure 2). That is to say, the combined action of hunting and carnivore predation on a common ungulate prey is additive and synergistic, not compensatory. So if Native American hunters even slightly lowered ungulate numbers, carnivores alone could continue to drive prey numbers lower, and keep those herds from recovering. Carnivore predation not only greatly complicates any harvest system, maximum sustained yield or otherwise, it also probably precluded Native Americans from developing specific practices to conserve ungulates.

Aboriginal Conservation

The American natives' role in nature has often been viewed as that of conservationists who were too wise and knowledgeable to overexploit their environment (McCabe and McCabe 1984:57). This belief, which can be traced to Rousseau's concept of the "noble savage," has a long history in the popular press (Speck 1913, 1939a, 1939b; Roberts 1932). The environmental movement in the late 1960s and early 1970s further romanticized the image of hunter-gatherers as original conservationists (Steinhart 1984). The idea that aboriginal peoples did not damage their resource base, which has been implicitly assumed by most anthropological theory since the beginning of that discipline (Heizer 1955), is still popular in some circles (Feit 1987, and others). I, however, have been unable to find any evidence that Native Americans effectively conserved ungulates.

Early writers claimed that defended exclusive use areas were com-



cumstances in which human harvest of moose triggered a catastrophic decline in both predator and prey. Without hunting, numbers of wolves, moose, and Dall sheep (*Ovis dalli*) are low but relatively stable. The addition of a small human harvest of moose, however, destabilizes the entire system. Even after hunting is halted, wolves continue to drive the moose population downward. The wolves then switch to Dall sheep and drive their numbers down as well. In this simulation, wolves go extinct before they can kill the few remaining ungulates, allowing prey populations to recover. This would not be the case, though, if humans continued to prey on the ungulates. Grizzily bear predation on newborn moose calves, and to a lesser extent on adults, is also important in this system, but that factor was not modeled separately. Instead, grizzily predation was included in calculation of moose survival rates internal to the model. Adapted from Haber (1977) and Walters et al. (1981).

mon, if not universal, among Native Americans (Heizer 1955:5). Although the origin of hunting territories has a long history of controversy, today most anthropologists hold the view that they were not aboriginal, but developed with the fur trade (e.g., Albers and Kay 1987; Bishop 1970). Even if we ignore the available evidence and grant that Native Americans were able to exclude other humans from their personal hunting territories, the problem is that they still never had exclusive use of the ungulates in those or any other areas. Carnivores were free to prey on any and all ungulates. This would have lowered territorial benefits and made exclusive use areas uneconomical to defend (Dyson-Hudson and Smith 1978). So it is likely that carnivore predation precluded the formation of hunting territories specifically designed to conserve ungulate numbers.

Aboriginal Overkill

Early writers also claimed that each native hunter "can tell at any time the number of animals which he can dispose of each year in his district without damaging his supply" (McLeod 1936:565). More recently, some anthropologists have asserted that native peoples harvested ungulates under principles similar to sustained yield management (Feit 1987; Nelson 1982), but Native American preference for prime-age females runs counter to any maximum sustained yield or conservation strategy.

Others have contended that Native Americans' religious belief systems prevented those peoples from overutilizing their resources (e.g., Speck 1939b; Nelson 1983). Native Americans tended to view wildlife as their spiritual kin and believed that success in the hunt was obtained by following prescribed rituals and atonement after the kill (e.g., Bettinger and Baumhoff 1982:503; Feit 1987). A scarcity of animals or hunting failures were not viewed as biological or ecological phenomena, but rather as a spiritual consequence of social events or circumstances. If a Native American could not find any game, it was not because he had overhar-vested the resource, but because he had done something to displease the gods. Since Native Americans saw no connection between their hunting and game numbers, their system of religious beliefs actually fostered the overexploitation of ungulate populations. Religious respect for animals does not equal conservation.

Jacobs (1971:237) and others have asserted that, "Native peoples lived according to what the conservationist Aldo Leopold has called a land ethic." Despite the widespread acceptance of the assumption that environmental attitudes predicate a society's actual environmental impacts, scholars have demonstrated the inadequacy of that paradigm (J. Kay 1985a, 1985b; Kay and Brown 1985; Tuan 1968, 1970). That is to say, there is no correlation between how a people say they are managing their resources and how they actually treat their environment. Belief systems are only a small part of the factors that influence how a people interact with their environment. There is no evidence that belief systems in and of themselves foster the actual implementation of conservation practices.

In summary, Native Americans had no practices that were specifically designed to conserve ungulates. All native hunters are essentially opportunistic and tend to take high-ranking ungulates regardless of the size of the prey populations or the likelihood of their becoming extinct. Native Americans had no concept of maximum sustained yield and did not manage ungulate populations to produce the greatest offtake. Human predation and predation by carnivores are additive and work in concert to reduce ungulate numbers. Moreover, competition from carnivores tended to negate any possible conservation practices.

ABORIGINAL POPULATIONS AND THE WILDERNESS MYTH

The idea that human populations purposefully limit their numbers so they will not overuse their resources has a long history in anthropology (Bates and Lees 1979). Some have claimed that the mere existence of human populations living in long-term relationships with ungulate species would seem, a priori, to argue for the existence of effective aboriginal management systems to prevent irrevocable depletion or extirpation of those resources (Freeman 1985). Coupled with this concept is the notion of limited needs. Hunter-gatherers, who were thought to need little, limited what they took from available resources (Hawkes et al. 1985).

enough for group selection to be an important evolutionary force, individual selection will always be more powerful (Heinen and Low 1992). considerations, not to preordained social restraints. Blurton-Jones cept of limited needs inaccurate; foragers respond instead to biologica constraints. Hawkes et al. (1985) and Hawkes (1987) have found the conselection argument. Since in practice groups do not go extinct often human population growth. strategies designed to maximize individual fitness may effectively limit Hawkes (1990, 1991, 1992, 1993) postulates that male reproductive tolerated-theft considerations, not to measures that conserve resources (1987a) suggests that low work rates may also be due to sharing or that birth intervals are limited by biological considerations, not by socia with resources. Blurton-Jones (1986, 1987b), however, has demonstrated though, have been cited as practices that balance human population The long intervals between births in some hunter-gatherer societies The idea that humans self-regulated their numbers relies on a grouprevisited, as well as the total pressure exerted on the resource base indicate how often resource patches or hunting areas would have been Questions regarding aboriginal numbers are important since they

Writers have long recognized that Native Americans lacked immunological resistance to epidemic and endemic European diseases and that many epidemics reduced aboriginal numbers by 50% to 90% at each passing (e.g., Cook and Lovell 1992; Stearn and Stearn 1945). Only recently, however, has it been realized that many epidemics swept in advance of even the earliest explorers. Dobyns (1983) postulated that Native American populations were severely reduced 100 to 200 years before the first European chroniclers. Ramenofsky (1987), who tested Dobyns's hypothesis against the archaeological record, found that the tribes along the middle Missouri River were decimated by European

disease ca. AD 1600, two hundred years before the arrival of Lewis and Clark. Campbell (1990) tested Dobyns's hypothesis against the archaeological record of the Columbia Plateau and concluded that European disease decimated those aboriginal populations ca. AD 1550. Taking this factor into consideration, several authors have recently revised aboriginal population estimates for North America upwards by as much as tenfold, to 100 million or more. These new figures suggest that precolumbian Native American populations were of sufficient size to make overexploitation of ungulates highly probable. Besides, as explained above (see Figure 2), human predation does not have to be particularly intense to trigger a collapse of ungulate numbers.

North America was not a "wilderness" waiting to be "discovered" but instead was home to tens of millions of aboriginal peoples before European-introduced diseases decimated their numbers. Prior to European arrival, most of this continent was owned, used, and modified by native peoples (e.g., Denevan 1992; Gomez-Pompa and Kaus 1992; Simms 1992). The idea that North America was a "wilderness" untouched by the hand of man prior to 1492 is a myth, a myth created, in part, to justify appropriation of aboriginal lands and the genocide that befell native peoples (Bowden 1992).

TESTING THE ABORIGINAL OVERKILL HYPOTHESIS

As discussed, my aboriginal overkill hypothesis developed from research in the Yellowstone ecosystem. To evaluate that paradigm's wider applicability, I tested it on several longstanding ecological problems in western North America. These included (1) why large herds of bison and other ungulates were absent from the Columbian Basin at historical contact and in precolumbian times; (2) the biogeography of moose in western North America—why moose were exceedingly rare or absent from the northern Rockies, most of British Columbia, and much of Alaska historically and prehistorically—areas that today support several hundred thousand moose; (3) the Kaibab deer incident in northern Arizona—what kept deer numbers low historically and prehistorically and why the population irrupted and severely overgrazed the range ca. 1920; and (4) why mule deer were rare in the Great Basin historically and in precolumbian times, only to irrupt in the early 1900s.

In each of these cases, I reviewed the various hypotheses that have been proposed to explain the observed variation in ungulate abundance. In all instances, I found that the available data do not support other

interpretations but do support the aboriginal overkill model. (Papers on each of these subjects are in preparation.) Thus, aboriginal overkill appears to be a robust hypothesis that applies not only to elk but also to moose, bison, mule deer, and other ungulates throughout the Intermountain West, and I suspect that it applies to other areas of the Americas as well. Birkedal (1993), for instance, suggested that aboriginal hunting even limited grizzly bear populations.

ARCHAEOLOGICAL DATA REVISITED

If my hypothesis is correct and if Dobyns's disease paradigm is true, the first wave of European diseases decimated Native American populations about 500 years ago, which in turn resulted in less hunting pressure and in greater numbers of ungulates, especially those most susceptible to overexploitation. These changes should then have been reflected by an increase in those species' relative abundances in human diets. As predicted, this pattern is observed in the archaeological record. Elk, in any number, first appeared in the intermountain archaeological record only 500 years ago (Frison 1991), and published reports indicate an identical pattern of mule deer abundance in Great Basin archaeological sites (Kay 1990). In Alaska and the western subarctic, moose bones do not appear in any numbers at archaeological sites until the past 500 years (Yesner 1989).

Similarly, if my aboriginal overkill hypothesis is correct, then archaeological sites used by aboriginal peoples who exploited or lived in cultural boundary zones should contain a higher proportion of elk remains than sites situated within the main cultural areas. And if Dobyns's disease hypothesis is correct, the proportion of elk should increase in those sites after ca. AD 1500. Such a pattern is observed, for instance, in the Nicola Valley of south-central British Columbia. "The Nicola Valley (and Canadian Okanogon) lack, to a great extent, the anadromous fish resources of the Thompson and Fraser Rivers to the west and north, and the American Okanogon to the south. This lack, combined with local topography, made [the Nicola] a valley region of comparatively low population density with social ties to the more resourceful region to the west and part of a prehistoric 'buffer zone' serving to maintain the Northern Plateau-Southern Plateau [cultural] boundary" (Wyatt 1972: abstract). At historical contact, an Athabaskan isolate, the Nicola, were surrounded by more numerous and powerful Interior Salish-speaking groups (Wyatt 1972:7).

Wyatt (1972) excavated eleven archaeological sites in the Nicola Val-

ley. Of the ungulate faunal remains recovered and dated between 2200 and 500 BP, elk represented 44%. Archaeological sites located in areas to the north and west (e.g., Langemann 1987; Rousseau and Richards 1988) and south (e.g., Campbell 1985) contain, on average, less than 5% elk. After ca. 500 BP, the proportion of elk in Nicola Valley sites increased to 70% of recovered ungulate remains (Wyatt 1972). So, spacial and temporal patterns of archaeologically recovered ungulate faunal remains support both the aboriginal overkill hypothesis and Dobyns's disease paradigm. There are, however, exceptions to aboriginal overkill.

EXCEPTIONS TO ABORIGINAL OVERKILL

According to predator-prey theory, prey populations will increase if they have a refugium where they are safe from predation (e.g., Taylor 1984). So, ungulates that could avoid aboriginal hunters should have been more abundant. Moreover, refugia do not have to be complete to be effective. Partial refugia will also enable prey populations to survive.

Unlike in other areas of the West, archaeological sites on the Washington, Oregon, and British Columbia coasts usually contain elk remains. Of the ungulate bones unearthed at those sites, elk constitute about 50% (Kay 1990). Thick coastal forests provided some refuge for elk because the plant communities were usually too wet to burn. Native peoples could not employ fire to open up the country and make hunting easier to the same extent that they did in other ecosystems (Lewis 1973, 1977; Lewis and Ferguson 1988). Because coastal regions receive little snowfall, aboriginal hunters also could not kill animals by chasing them into deep snow as natives commonly did elsewhere.

Early explorers reported that elk were also common in California's Central Valley along the Sacramento and San Joaquin rivers (McCullough 1971). When disturbed, however, those animals would thee into swamps where they could not be hunted. This behavior was observed by John Work (1945:62), who led a Hudson's Bay Company fur brigade through California in 1831–1832.

The people are rather short of food and no more can be got, the hunters are not able to kill the elk. There are a good many along the marshy borders of the lake but they seldom venture out on the hard ground and when any of them happen to be found out, they fly immediately in among the water and bulrushes where they cannot be pursued.

Work noted, however, that when floodwaters forced elk from the swamps, they were easily killed by native peoples, who often simply ran

the animals down and killed them with knives or spears. Without refuge provided by the tule swamps, large herds of elk would not have survived in California's Central Valley.

Herds of bison on the Great Plains and caribou in the Arctic had no physical refugia; instead they had refugia in time. By undertaking extensive migrations, bison and caribou were able to outdistance most of their human and carnivorous predators. Wolves with young, for instance, simply could not keep pace with or even follow the migrating herds (Bergerud 1990, 1992). The same was true of humans who had to transport children, as well as their possessions. Caribou that migrate long distances today have densities ten times greater than nonmigratory populations (Seip 1991).

Research in Africa's Serengeti has shown that resident ungulates are limited by predators while migratory animals are not and that Serengeti ungulates migrate primarily to avoid predation, not to secure food (Crete and Huot 1993:2295; Fryxell et al. 1988). I suggest that this was also the case in North America, and that bison and caribou would have been much less abundant if they had not migrated long distances. Migration not only took bison and caribou beyond the reach of most humans, but the Great Plains and the Arctic tundra provided few alternative foods that could sustain aboriginal populations when ungulates migrated. As noted above, tribal boundary or buffer zones also provided refugia for ungulate populations. Without refugia, few ungulates would have been able to withstand the onslaught of human predators.

IMPLICATIONS OF ABORIGINAL OVERKILL

My ideas regarding predation by Native Americans have significant implications for conservation biology, management of natural areas, wilderness management, national park management, range management, and wildlife biology, as well as anthropology-archaeology, since those disciplines seldom consider the impact prehistoric human populations had on their resource base or how aboriginal activities may have structured entire ecosystems. For instance, most national parks, wilderness areas, and natural areas are supposedly managed to represent the conditions that existed in precolumbian times (i.e., so-called natural or pristine conditions). But what is natural? If Native Americans limited ungulate numbers, which in turn determined the structure of entire plant and animal communities, that is a completely different situation than letting resource-limited ungulates do the same (Wagner and Kay 1993). A "hands-off" or "natural regulation" approach by modern land

managers will not duplicate the ecological conditions under which those communities developed. If aboriginal predation and burning created those communities, then the only way to maintain what we call "natural areas" today is to duplicate aboriginal influences and processes.

Systems with native peoples are entirely different from those without aboriginal populations (e.g., Western and Gichohi 1993). In fact, the modern concept of wilderness, as areas without human influence, is a myth. As Gomez-Pompa and Kaus (1992) have pointed out, the only "wilderness" is in the mind of Europeans. Setting aside an area as wilderness today will not preserve some remnant of the past but instead create conditions that have not existed for the past 10,000 years. Here in North America, for instance, we view the Amazon as a wilderness to be saved and protected, but to indigenous peoples it is a home—a home they have modified to suit human needs (e.g., Balée 1989).

strategy (e.g., Butzer 1992; Denevan 1992; Diamond 1988, 1992; Heinen strategy. For humans, conservation is seldom an evolutionarily stable was, in most instances, the exact opposite of any predicted conservation ungulates, but the manner in which those peoples harvested ungulates not only did Native Americans have no effective practices to conserve maximum sustained yield. The data I have collected, though, show that did not overuse their resources and harvested ungulate populations at Likewise, anthropologists have formulated many of their ideas of pre-historic human ecology around the assumption that aboriginal peoples once structured entire ecosystems (Mills et al. 1993). What I am proposronment. Native Americans were the ultimate keystone species that mized their individual fitness regardless of their impacts on the envioverexploit their resources, Native Americans acted in ways that maxithey actually do. Instead of being "noble savages" who were too wise to and Low 1992). I have also been unable to find any evidence that Native should be viewed between how a people say they are managing their resources and what from overutilizing ungulate populations. There is little correlation Americans' system of religious beliefs prevented aboriginal peoples ing is a major paradigm shift of how ecosystems and aboriginal peoples

I am indebted for helpful comments from Fred Wagner, Jim O'Cornell, Kristen Hawkes, Robert Taylor, Steve Simms, Paul Martin, and Henry Dobyns, among others. My research in Yellowstone was funded by the Welder Wildlife Foundation and Utah State University's Ecology Center. My Aboriginal Overkill project has been supported by USU's Institute of Political Economy and its Ecology Center. Jim O'Connell, Kristen Hawkes, Steve Simms, Michael Alvard, Randy Simmons, Robert Taylor, and two anonymous reviewers read drafts of this paper and offered suggestions that materially improved its content.

stone ecosystem is in press and another book on aboriginal overkill is forthcoming. ecosystem states and processes in the central Canadian Rockies. A book on the Yellow-University. He is presently working for Parks Canada on an assessment of the long-term is a Natural Resource Policy Associate with the Institute of Political Economy at Utah State The author received his Ph.D. in wildlife ecology from Utah State University in 1990 and

NOTES

- cluded developing many topics in detail. overkill hypothesis, not the full weight of available evidence. Space also preseveral thousand citations. This paper contains only an outline of my aboriginal November 1993. A forthcoming book will include additional data, as well as 1. This paper is based on a presentation given at the 92nd annual meeting of the American Anthropological Association held at Washington, D.C. in
- able to aboriginal people. This assumption, though, is incorrect (as shown in this Human subsistence studies invariably make this same assumption: that is, carnivore predators had little or no impact on the number of ungulates avail-
- elk were rare where they are now common. nal remains recovered from archaeological sites in the southern Canadian Rock- Kay (1990) contains MNI and NISP data for ungulates unearthed at near-ly 300 individual archaeological sites in Wyoming, Montana, Utah, Washington, Oregon, Nevada, and Idaho. I have also completed an analysis of ungulate fauies, and those data exhibit a similar pattern (Kay et al. 1994). Archaeologically,
- and firearms. So, native participation in the fur trade cannot be blamed for the was one of the last regions to be explored in the western United States (Haines absence of game in Yellowstone ca. 1800, as was true in some areas.

 5. Although my research has been limited to western North America, 1 Shoshone, who inhabited Yellowstone at historical contact, lacked both horses was subject to less direct European disturbance than other areas. The Mountain 1977). Moreover, Yellowstone was not on established travel routes and therefore 4. Yellowstone was not "officially" discovered by Europeans until 1869 and
- suspect aboriginal overkill may be a universal attribute of hunter-gatherer societies, except where prey (ungulates) have refugia (discussed in the section on "Exceptions").
- ranging ungulate population not subject to human exploitation (e.g., Hamlin can neither increase or decrease. These conditions rarely, if ever, occur in the and Mackie 1989). wild, and there is no evidence that all three occur simultaneously in any freebirth rate, constant age-specific death rates, and a constant size-the population These data will be presented in forthcoming publications.

 To exhibit a stable age distribution, a population must have a constant
- assume that carnivore predation was unimportant and that humans harvested nivores played in limiting prehistoric ungulate populations, which, in turn, constrained human subsistence patterns. Invariably, human subsistence studies late population densities (e.g., Osborn 1993). Since both these assumptions are incorrect, published human subsistence strategies must be viewed with caution. populations at maximum sustained yield based on modern food-limited unguanthropologist or archaeologist who has seriously considered the role that car-8. Aside from Mithen (1986, 1987, 1989, 1990), I am not aware of any

Aboriginal Overkill

- lations to prosper. densities, as previously discussed, was an artifact of group or tribal interacriginal buffer zones cannot be considered a conservation practice per se (Alvard tions, not the outcome of practices designed to conserve ungulates. Thus, abo-1993a, 1993b) even though they may have permitted localized ungulate popu-The formation of aboriginal buffer zones with their higher ungulate
- 10. The exact impact of native hunting varied, depending on the abundance of alternative human foods, the behavior of the specific ungulates, the techniques used to hunt those ungulates, and whether or not the ungulates had refugia where they could escape predation, among other factors

REFERENCES

Albers, P., and J. Kay

1987 Sharing the Land: A Study in American Indian Territoriality. In A Cul-Pp. 47-91. Boulder, Colorado: Westview Press. tural Geography of North American Indians, T. E. Ross and T.G. Moore, eds

Alvard, M. S.

1993a Testing the Ecological Noble Savage Hypothesis: Conservation and Subsistence Mexico, Albuquerque. Hunting by the Piro of Amazonian Peru. Ph.D. Dissertation, University of New

Anderson, A. E., D. E. Medin, and D. C. Bowden 1993b Testing the "Ecologically Noble Savage" Hypothesis: Interspecific Prey Choice by Piro Hunters of Amazonian Peru. Human Ecology 21:355-387.

Anderson, M. K. 1972 Indices of Carcass Fat in a Colorado Mule Deer Population. Journal of Wildlife Management 36:579-594.

1991 Plant Gathering as a Conservation Strategy: Learning from California's Natural Areas Association Seventeenth Annual Meeting Future, J. Edelborck and S. Carpenter, eds. Pp. 472-481. Proceedings of the Earliest Resource Managers. In Natural Areas and Yosemite: Prospects for the

1969 Running Down and Driving of Game in North America. Studia Etlinographica Upsaliensia 30:1–129.

Balée, W.

Bates, D. G., and S. H. Lees 1989 The Culture of Amazonian Forests. Advances in Economic Botany 7:1-21.

1979 The Myth of Population Regulation. In Evolutionary Biology and Human ate, Massachusetts: Duxbury Press Social Beliavior, N. A. Chagnon and W. Irons, eds. Pp. 273-289. North Scitu

Bergerud, A. T.

1983 Prey Switching in a Simple Ecosystem. Scientific American 249:130-136

1990 Rareness as an Anti-predator Strategy to Reduce Predation Risk. In Popceedings Held September 1989, Trondheim, Norway, Vol. 1. ulation Dynamics, pp. 15-25. Transactions of 19th I.U.G.B. Congress: Pro

1992 Rareness as an Antipredator Strategy to Reduce Predation Risk for R. Barrett, eds. Pp. 1008-1021. New York: Elsevier Applied Science Moose and Caribou. In Wildlife 2001: Populations, D. M. McCullough and

Bettinger, R. L., and M. A. Baumhoff

1982 The Numic Spread: Great Basin Cultures in Competition. American Antiq uity 47:485-503

Binford, L. R.

1978 Nunamiut Ethnoarchaeology. New York: Academic Press

1981 Bones: Ancient Men and Modern Myths. New York: Academic Press

1993 Ancient Hunters in the Alaskan Wilderness: Human Predators and Their and S.D. Veirs, Jr., eds. Pp. 228–234. Hancock, Michigan: The George Wright Management. In Partners in Stewardship: Proceedings of the 7th Conference on Role and Effect on Wildlife Populations and the Implications for Resource Research and Resource Management in Parks and on Public Lands, W. E. Brown

Bishop, C. A.

1970 The Emergence of Hunting Territories among the Northern Ojibwa. Eth nology 9:1-15.

Blurton-Jones, N. G.

1986 Bushman Birth Spacing: A Test for Optimal Interbirth Intervals. Ellmology and Sociobiology 7:91-105.

1987a Tolerated Theft, Suggestions about the Ecology and Evaluation of Sharing, Hoarding and Scrounging. Social Science Information 26:31-54.

1987b Bushman Birth Spacing: Direct Tests of Some Simple Predictions. Ethnology and Sociobiology 8:183-203

Bowden, M. J.

1992 The Invention of American Tradition. Journal of Historical Geography 18:3-26.

Boyce, M. S.

1991 Natural Regulation or the Control of Nature? In The Greater Yellowstone Ecosystem: Redefining America's Wilderness Heritage, R. B. Keiter and M. S. Boyce, eds. Pp. 183-208. New Haven: Yale University Press.

Butzer, K. W.

Campbell, S. K. 1992 The Americas before and after 1492: An Introduction to Current Geo graphical Research. Association of American Geographers Annals 82:345-368

1990 Post Columbian Cultural History in Northern Columbia Platcau, A.D. 1500-1900. New York: Garland Publishing.

Campbell, S. K., ed.

1985 Summary of Results, Chief Joseph Cultural Resources Project, Washington. Office of Public Archaeology, University of Washington, Seattle.

Carbyn, L. N., S. M. Oosenbrug, and D. W. Anions

1993 Wolves, Bison, and the Dynamics Related to the Peace-Athabasca Delta in sity of Alberta, Edmonton. Canada's Wood Buffalo National Park. Circumpolar Research Series 4. Univer-

Aboriginal Overkil

1984 The Energetic Paradox of Human Running and Hominid Evolution. Current Anthropology 25:483-493.

Caughley, G.

1976 Wildlife Management and the Dynamics of Ungulate Populations. Applied Biology 1:183-246.

Chadde, S. W., and C. E. Kay

1988 Willows and Moose: A Study of Grazing Pressure, Slough Creek Exclosure, iment Station Research Note 24. Montana, 1961-1986. University of Montana Forest and Conservation Exper-

1991 Tall Willow Communities on Yellowstone's Northern Range: A Test of the "Natural Regulation" Paradigm. In The Greater Yellowstone Ecosystem: Pp. 231–264. New Haven: Yale University Press Redefining American's Wilderness Heritage, R. R. Keiter and M.S. Boyce, eds

Chamberlin, R. V.

1911 The Ethnobotany of the Gosiute Indians of Utah. Memors of the American Anthropological Association 2(5):331-405.

Clutton-Brock, T. H., M. Major, and F. E. Guinness

1985 Population Regulation in Male and Female Red Deer. Journal of Animal Ecology 54:831-846.

Cohen, M. N.

1977 The Food Crisis in Prehistory. New Haven: Yale University Press

Cook, N. D., and W. G. Lovell, eds.

1992 Secret Judgements of God: Old World Disease in Colonial Spanish America. Norman: University of Oklahoma Press.

Coughenour, M. B., and F. J. Singer

1991 The Concept of Overgrazing and Its Application to Yellowstone's Northness Heritage, R. B. Keiter and M. S. Boyce, eds. Pp. 209-230. New Haven: em Range. In The Greater Yellowstone Ecosystem: Redefining America's Wilder-Yale University Press.

Crete, M., and J. Huot

1993 Regulation of a Large Herd of Migratory Caribou: Summer Nutrition gy 71:2291-2296. Affects Calf Growth and Body Reserves of Dams. Canadian Journal of Zoolo-

Denevan, W.

1992 The Pristine Myth: The Landscape of the Americas in 1492. Association of American Geographers Annals 82:369-385

Despain, D., D. Houston, M. Meagher, and P. Schullery

1986 Wildlife in Transition: Man and Nature on Yellowstone's Northern Range. Boulder, Colorado: Roberts Rínehart.

1988 The Golden Age That Never Was. Discover 9(12):70-79

1992 The Third Chimpanzee: The Evolution and Future of the Human Animal. New York: Harper Collins.

Dobyns, H. F.

1983 Their Numbers Become Thinned: Native American Population Dynamics in Eastern North America. Knoxville: University of Tennessee Press

Dyson-Hudson, R., and E. A. Smith

1978 Human Territoriality: An Ecological Reassessment. American Anthropolo gist 80:21-41.

Eberhardt, L. L.

1987 Population Projections from Simple Models. Journal of Applied Ecology

Flook, D. R. 1987 North American Native Hunting and Management of Moose Populations. Swedish Wildlife Research (Supplement 1):25–42.

1970 A Study of Sex Differential in the Survival of Wapiti. Canadian Wildlife

Service Report Series 11:1-71.

Freeman, M. M. R. 1985 Appeal to Tradition: Different Perspectives on Arctic Wildlife Manage ples, J. Brosted et al., eds. Pp. 265-281. Universitetsforlagetus Bergen. ment. Native Power: The Quest for Autonomy and Nationhood of Indigenous Peo-

Frison, G. C.

1971 Shoshonean Antelope Procurement in the Upper Green River Basin, Wyoming. Plains Anthropologist 16:258-284.

1991 Prehistoric Hunters of the High Plains, second ed. New York: Academic Press.

Fritts, S. H., and L. D. Mech Frison, G. C., and D. N. Walker, eds.

1984 The Dead Indian Creek Site: An Archaic Occupation in the Absaroka Mountains of Northeastern Wyoming. Wyoming Archaeologist 27(1-2):11–122.

1981 Dynamics, Movements, and Feeding Ecology of a Newly Protected Wolf Population in Northwestern Minnesota. Wildlife Monographs 80.

Fryxell, J. M., J. Greever, and A. R. E. Sinclair

1988 Why Are Migratory Ungulates So Abundant? American Naturalist 131 781-798.

Fuller, T. K., and L. B. Keith

1980 Wolf Population Dynamics and Prey Relationships in Northeastern Alberta. Journal of Wildlife Management 44:583-602.

Stephenson, and D. G. Larsen Gasaway, W. C., R. D. Boertje, D. V. Grangaard, D. G. Kellyhouse, R. O.

and Implications for Conservation. Wildlife Monographs 120. Gasaway, W. C., R. O. Stephenson, and J. L. Davis 1992 The Role of Predation in Limiting Moose at Low Densities in Alaska and Yukon

1983 Wolf-Prey Relationships in Interior Alaska. Alaska Department of Fish and Game Wildlife Technical Bulletin 6.

1986 New Evidence of High Frequency Antler Wounding in Cervids. Canadian Journal of Zoology 64:380-384

Gomez-Pompa, A., and A. Kaus

1992 Taming the Wilderness Myth. Bioscience 42:271-279

Aboriginal Overkill

1982 To Have or Have Not: The Ecology of Sharing among Hunter-gatherers. In Resource Managers: North American and Australian Hunter-gatherers, N. M. Boulder, Colorado: Westview Press. Williams and E. S. Hunn, eds. Pp. 69-92. AAAS Selected Symposium 67

1977 Socioecological Dynamics of Wolces and Prey in a Subarctic Ecosystem. Ph.D. Dissertation, University of British Columbia, Vancouver.

1965 Valley of the Upper Yellowstone. Norman: University of Oklahoma Press. 1977 The Yellowstone Story. Vol. 1. Yellowstone Library and Museum Associastone National Park, Wyoming. tion in cooperation with Colorado Association University Press, Yellow

Hamlin, K. L., and R. J. Mackie

1989 Mule Deer in the Missouri River Breaks, Montana: A Study of Population Dynamics in a Fluctuating Environment. Helena: Montana Department of Fish, Wildlife, and Parks.

Harris, A.

1978 The Mummy Cave Tetropods. In The Mummy Cave Project in Northwestern Wyoming, H. McCracken, ed. Pp. 146–151. Cody, Wyoming: Buffalo Bill Historical Center.

Hastings, A.

1983 Age-dependent Predation Is Not a Simple Process, I: Continuous Time Models. Theoretical Population Biology 23:347-362.

1984 Age-dependent Predation Is Not a Simple Process, II: Wolves, Ungulates, and a Discrete Time Model for Predation on Juveniles with a Stabilizing Tail. Theoretical Population Biology 26:271-282.

1987 How Much Food Do Foragers Need? In Food and Evolution, M. Harris and E. B. Ross, eds. Pp. 341–355. Philadelphia: Temple University Press.

1990 Why Do Men Hunt? Some Benefits for Risky Strategists. In Risk and Boulder, Colorado: Westview Press. Uncertainty in Tribal and Peasant Economics. E. Cashdan, ed. Pp. 145-166

1991 Showing off: Tests of an Hypothesis about Men's Foraging Goals. Ethol ogy and Sociobiology 12:29-54.

1992 On Sharing and Work. Current Anthropology 33:401-407.

1993 Why Hunter-gatherers Work. Current Anthropology 34:341-361

Hawkes, K., and J. F. O'Connell

1981 Affluent Hunters? Some Comments in Light of the Alyawara Case. Amer. ican Anthropologist 83:622-626.

Hawkes, K., J. F. O'Connell, K. Hill, and E. L. Charnov

1985 How Much Is Enough? Hunters and Limited Needs. Ethnology and Socio biology 6:3-15.

Heinen, J., and B. Low

1992 Human Behavioral Ecology and Environmental Conservation. Environ mental Conservation 19:105-116.

1955 Primitive Man as an Ecological Factor. Kroeber Anthropological Society Paper 13:1-31.

1965 The Virginia Deer and Intertribal Buffer Zones in the Upper Mississipp Adjustments, A. Leeds and A. P. Vayda, eds. Pp. 43-65. American Associa-Valley. In Man, Culture and Animals: The Role of Animals in Human Ecological tion for the Advancement of Science Publication 78.

Hoskinson, R. L., and L. D. Mech 1976 White-tail Deer Migration and Its Role in Wolf Predation. Journal of Wildlife Management 40:429-441

Houston, D. B.

Jacobs, W. R. 1982 The Northern Yellowstone Elk: Ecology and Management. New York: MacMillan

1971 The Fatal Confrontation: Early Native-White Relations on the Frontiers tory Review 40:283-309. of Australia, New Guinea, and America—A Comparative Study. Pacific His

Johns, P. E., M. H. Smith, and R. K. Chesser

1984 Annual Cycles of the Kidney Fat Index in a Southeastern White-tailed Deer Herd. Journal of Wildlife Management 48:969-973.

1985 Aspen Reproduction in the Yellowstone Park-Jackson Hole Area and Its Relationship to the Natural Regulation of Ungulates. In Western Elk Management: A Symposium, G. W. Workman, ed. Pp. 131-160. Logan: Utah State

1987 Too Many Elk in Yellowstone? Western Wildlands 13(3):39-41, 44.

1990 Yellowstone's Northern Elk Herd: A Critical Evaluation of the "Natural Regulation" Paradigm. Ph.D. Dissertation, Utah State University, Logan.

1992a An Alternative Interpretation of the Historical Evidence Relating to the Management, August 25-27, 1992. University of Alberta, Edmonton. Second North American Symposium on Wolves: Their Status, Biology, and Abundance of Wolves in the Yellowstone Ecosystem. Paper presented at the

1992b Book Review-The Jackson Hole Elk Herd: Intensive Wildlife Management

in North America. Journal of Range Management 45:315-316.
1993a Aspen Seedlings in Recently Burned Areas in Grand Teton and Yel lowstone National Parks. Northwest Science 67:94-104.

1993b Effects of Browsing by Native Ungulates on Shrub Growth and Seed ber 19-21, 1993. Las Vegas. ed at the Symposium on Wildland Shrub and Arid Land Restoration, Octotation, Restoration, and "Natural Regulation" Management. Paper present-Production in the Greater Yellowstone Ecosystem: Implications for Revege-

1994 The Impact of Native Ungulates and Beaver on Riparian Communities in the Intermountain West. Natural Resources and Environmental Issues 1:23-44.

Kay, C. E., and S. W. Chadde

1992 Reduction of Willow Seed Production by Ungulate Browsing in Yellowof Riparian Shrub Communities, W. P. Clary, E. D. McArthur, D. Bedunah stone National Park. In Proceedings-Symposium on Ecology and Management

> Report INT-289. and C. L. Wambolt, eds. Pp. 92-99. U.S. Forest Service General Technical

Kay, C. E., and F. H. Wagner

1991 Historic Condition of Woody Vegetation on Yellowstone's Northern lowstone National Park, Mammoth, Wyoming. sented at "Plants and Their Environments-First Biennial Scientific Confer-Range: A Critical Test of the "Natural Regulation" Paradigm. Paper preence on the Greater Yellowstone Ecosystem," September 16-17, 1991. Yel

Kay, C. E., B. Patton, and C. A. White

1994 Assessment of Long-term Terrestrial Ecosystem States and Processes in Parks Canada, Banff National Park, Alberta. Banff National Park and the Central Canadian Rockies. Report prepared for

1985a Native Americans in the Fur Trade and Wildlife Depletion. Environmental Review 9:118-130.

1985b Preconditions of Natural Resource Conservation. Agricultural History 59:124-135.

Kay, J., and C. J. Brown

1985 Mormon Beliefs about Land and Natural Resources, 1847-1877. Journal of Historical Geography 11:254-267

1987 Zooarchaeology of the Lillooet Region, British Columbia. M.A. Thesis, Simon Fraser University, Burnaby, British Columbia.

Langford, N. P.

1972 The Discovery of Yellowstone Park. Lincoln: University of Nebraska Press

Lee, R. B., and I. DeVore, eds. 1968 What Hunters Do for a Living, or How to Make Out on Scarce Resources. In Man the Hunter, R. B. Lee and I. DeVore, eds. Pp. 30-48. Chicago: Aldine.

1968 Man the Hunter. Chicago: Aldine

Lewis, H. T.

1973 Patterns of Indian Burning in California: Ecology and Ethnohistory. Anthropological Papers No. 1. Pamona, California: Ballena Press.

1977 Maskuta: The Ecology of Indian Fires in Northern Alberta, Wistern Canadian Journal of Anthropology 7:15-52.

Lewis, H. T., and T. A. Ferguson

1988 Yards, Corridors and Mosaics: How to Burn a Boreal Forest. Human Ecolогу 16:57-77.

Lewis, M., and W. Clark

1893 The History of the Lewis and Clark Expedition, 3 vols. E. Coues, ed. New York: Francis P. Harper. Republished by Dover Publications, New York.

1924 Notes on Shoshoni Ethnography. American Museum of Natural History Anthropological Papers 20(3):185-314.

Lowie, R. H.

1987 On the Analysis of Vertebrate Mortality Profiles: Sample Size, Mortality Type and Hunting Pressure. American Antiquity 52:125-142

McCabe, R. E., and T. R. McCabe

1984 Of Slings and Arrows: A Historical Retrospection. In White-tailed Deer. Management Institute and Stackpole Books Ecology and Management, L. K. Halls, ed. Pp. 19-72. Harrisburg: Wildlife

1971 The Tule Elk: Its History, Behavior, and Ecology. University of California Publications in Zoology 88.

McLeod, W. C.

Mech, L. D. 1936 Conservation among Primitive Peoples. Scientific Monthly 43:562-566

1977 Wolf-pack Buffer Zones as Prey Reservoirs. Science 198:320-321.

1994 Buffer Zones of Territories of Gray Wolves as Regions of Intraspecific Strife. Journal of Mammalogy 75:199-202.

1985 Social Organization, Spatial Distribution, and Population Density of Wolves in Relation to Moose Density. Canadian Journal of Zoology 63:

1991 The Significance of Limiting and Regulating Factors on the Demography of Moose and White-tailed Deer. Journal of Animal Ecology 60:377-393

1994 Ungulate Population Models with Predation: A Case Study with the North American Moose. Ecology 75:478-488.

Messier, F., and M. Crete

1985 Moose-Wolf Dynamics and the Natural Regulation of Moose Popula tions. Œcology 65:503-512.

Michod, R. E.

1979 Evolution of Life Histories in Response to Age-specific Mortality Factors American Naturalist 113:531-550.

Mills, L. S., M. E. Soule, and D. F. Doak

1993 The Keystone-species Concept in Ecology and Conservation. Bioscience 43:219-224.

Mithen, S. J.

1986 Reindeer and Risk: Simulating Upper Palaeolithic Economies in the Perigord. Archaeological Review from Cambridge 5:88-105

1987 Modelling Decision Making by Low-latitude Hunter Gatherers. Journal of Operations Research 30:240-242

1989 Modeling Hunter-gatherer Decision Making: Complementing Optima Foraging Theory. Human Ecology 17:59-83

1990 Thoughtful Foragers. New York: Cambridge University Press

Nelson, L. J., and J. M. Peek

1982 Effect of Survival and Fecundity on Rate Increase of Elk. Journal of Wildlife Management 46:535-540.

Nelson, R. K.

1973 Hunters of the Northern Forest: Designs for Survival among the Alaskan Kutchin. Chicago: University of Chicago Press

1982 A Conservation Ethic and Environment: The Koyukon of Alaska. In Resource Managers: North America and Australian Hunter-gatherers, N. M.

> Boulder, Colorado: Westvicw Press. Williams and E. S. Hunn, eds. Pp. 211-228. AAAS Selected Symposium 67

1983 Make Prayers to the Russen. Chicago: University of Chicago Press

Nimmo, B. W.

1971 Population Dynamics of a Wyoming Pronghorn Cohort from the Eden-Farson Site, 48SW304. Plains Anthropologist 16:285-288

O'Connell, J. F., K. Hawkes, and N. Blurton-Jones

Okarına, H. 1988 Hadza Hunting, Butchering, and Bone Transport and Their Archaeolog-1984 The Physical Condition of Red Deer Falling Prey to the Wolf and Lynx ical Implications. Journal of Anthropological Research 44:113-161

and Harvested in the Carpathian Mountains. ACTA Therlogica 29:283-290.

Osborn, A. J.

1993 Snowblind in the Desert Southwest: Moisture Islands, Ungulate Ecology. ical Research 49:135-164. and Alternative Prehistoric Overwintering Strategies. Journal of Anthropolog-

Patten, D. T.

1993 Herbivore Optimization and Overcompensation: Does Native Herbivory 3:35-36.on Western Rangelands Support These Theories? Ecological Applications

Peterson, R. O.

1977 Wolf Ecology and Prey Relationships on Isle Royale. National Park Service Science Monograph 11.

Post, R. H.

1938 The Sinkaietk or Southern Okanayon of Washington: The Subsistence Quest. In The Sinkaietk or Southern Okanagon of Washington, L. Spier, ed. Pp. 11-34. General Series in Anthropology No. 6. Menasha, Wisconsin: George Banta Publishing.

Ramenofsky, A. F.

1987 Vectors of Death: The Archaeology of European Contact. Albuquerque: University of New Mexico Press.

Raynolds, F. W.

1868 Report on the Exploration of the Yellowstone River in 1859-60. Senate Executive Document 77, 40th Congress, 2nd Session.

1932 Conservation as Formerly Practiced by the Klamath River Region. California Fish and Game 18:283–290.

Roper, E. 1986 Steamboat Elk-Carter Leases Mule Deer-Subletle Antelope Project Quarterly Progress Report, November 1, 1985-February 15, 1986. Wyoming

Rousseau, M. K., and T. H. Richards Game and Fish Department, Green River, Wyoming

1988 The Oregon Jack Creek Site (EdRi-6): A Lehman Phase Site in the Thompson River Valley, British Columbia. Canadian Journal of Archaeology

1972 Stone Agc Economics. Chicago: Aldine Press

1991 Predation and Caribou Populations. Rangifer (Special Issue) 7:46-52

1992 Factors Limiting Woodland Caribou Populations and Their Interrelation Journal of Zoology 70:1494-1503. ships with Wolves and Moose in Southeastern British Columbia. Canadian

Short, H. L.

1979 Decr in Arizona and New Mexico: Their Ecology and a Theory Explaining Recent Population Decreases, U.S. Forest Service General Technical Report RM-70

Simms, S. R.

1984 Aboriginal Great Basin Foraging Strategies: An Evolutionary Analysis. Ph.D Dissertation, University of Utah, Salt Lake City.

1992 Wilderness as a Human Landscape. In Wilderness Tapestry, S. I. Zeveloff, L. M. Vause, and W. H. McVaugh, eds. Pp. 183-201. Reno: University of

Simpson, T.

Sinopoli, C. M. 1984 Population Dynamics of Mule Deer. Wyoming Archaeologist 27(1-2):83-96

1991 Style in Arrows: A Study of an Ethnographic Collection from the West-Studies 10:63-87 ern United States. Michigan Discussions in Anthropology, Hunter-Gathere

Smith, A. M.

1974 Ethnography of the Northern Utes. Museum of New Mexico Paper in Anthropology 17

1983 Anthropological Applications of Optimal Foraging Theory: A Critical Review. Current Anthropology 24:625-651.

1991 Inuijuamiut Foraging Strategies: Evolutionary Ecology of an Arctic Hunting Economy. New York: Aldine de Gruyter.

Smith, E. A., and B. Winterhalder, eds.

Speck, F. G. 1992 Evolutionary Ecology and Human Beliavior. New York: Aldine de Gruyter

1913 The Indians and Game Preservation. Red Man 6:21-25

1939a Savage Savers. Fronticrs 4:23-27

1939b Aboriginal Conservators. Bird Lore 40:258-261

Speth, J. D.

1983 Bison Kills and Bone Counts: Decision Making by Ancient Hunters. Chicago: University of Chicago Press.

1987 Early Hominid Subsistence Strategies in Seasonal Habitats. Journal of Archaeological Science 14:13-29.

1989 Early Hominid Hunting and Scavenging: The Role of Meat as an Energy Source. Journal of Human Evolution 18:329-343.

1991 Protein Selection and Avoidance Strategies of Contemporary and Ances ety of London (Series B) 334:265-270 tral Foragers: Unresolved Issues. Philosophical Transactions of the Royal Seci-

Aboriginal Overkill

Speth, J. D., and K. A. Spielmann 1983 Energy Source, Protein Metabolism, and Hunter-gatherer Subsistence Strategies. Journal of Anthropological Archaeology 2:1-31

Spielmann, K. A.

1989 A Review: Dietary Restrictions on Hunter-gatherer Women and the Implications for Fertility and Infant Mortality. Human Ecology 17:321-345.

Stearn, E. W., and A. E. Stearn

1945 The Effect of Smallpox on the Destiny of the Amerindian. Boston: Bruce Humphries.

1984 Ecological Saints. Auduban 86(4):8-9

Stephens, D. W., and J. R. Krebs

1986 Foraging Theory. Princeton: Princeton University Press

Taylor, R. J.

1984 Predation. New York: Chapman and Hill

Teit, J. H.

1928 The Middle Columbia Salish. University of Washington Publications in Anthropology 2:83-128

Temple, S. A.

1987 Do Predators Always Capture Substandard Individuals Disproportionately from Prey Populations? Ecology 68:669-674

Townsend, J. K.

1978 Narrative of a Journey across the Rocky Mountains to the Columbia River (Originally published in 1839). Lincoln: University of Nebraska Press.

1968 Discrepancies between Environmental Attitude and Behavior: Examples from Europe and China. Canadian Geographer 12:176-190.

1970 Treatment of the Environment in Ideal and Actuality. American Scientis 58:244-249.

Voigt, D. R., G. B. Kolenosky, and D. H. Pimlott 1976 Changes in Summer Foods of Wolves in Central Ontario. Journal of Wildlife Management 40:663-668.

Voorhies, M. R.

1969 Taphonomy and Population Dynamics of an Early Pliocene Vertebrate Fauna, Knox County, Nebraska. University of Wyoming Contribution to Geology, Special Paper 1:1-69.

Wagner, F. H., and C. E. Kay

1993 "Natural" or "Healthy" Ecosystems: Are U.S. National Parks Providing Them? In Humans as Components of Ecosystems, M. J. McDonnell and S. T. Pickett, eds. Pp. 257-270. New York: Springer-Verlag.

Walters, C. J., M. Stocker, and G. C. Haber

1981 Simulation and Optimization Models for a Wolf-Ungulate System. In Pp. 317-337. New York: John Wiley and Sons Dynamics of Large Mammal Populations, C. W. Fowler and T. D. Snith, eds

Washburn, S. L., and C. S. Lancaster

1968 The Evolution of Hunting. In Man the Hunter, R. B. Lee and I. DeVore, eds. Pp. 392-303. Chicago: Aldine

Some

Webster, G. S.

1979 Deer Hides and Tribal Confederations: An Appraisal of Gramly's Hypothesis. American Antiquity 44:816-820.

1983 Iroquoian Hunting: An Optimization Approach. Ph.D. Dissertation, Pennsylvania State University, State College.

Western, D., and H. Gichohi

for Ecosystem Viability Analysis. African Journal of Ecology 31:269–281. Winterhalder, B. 1993 Segregation Effects and the Impoverishment of Savanna Parks: The Case

Work, J. 1987 An Analysis of Hunter-gatherer Diets: Stalking an Optimal Foraging Model. In Food and Evolution: Toward a Theory of Human Fcod Habits, M. Harris and E. B. Ross, eds. Pp. 311-339. Philadelphia: Temple University Press

1945 Fur Brigade to the Bonaventura: John Work's California Expedition, 1832-1833 for the Hudson's Bay Company, A. B. Maloney, ed. San Francisco: California Historical Society.

Wright, G. A.

1984 People of the High Country: Jackson Hole before the Settlers. New York: Peter

Wright, G. A., and S. J. Miller
1976 Prehistoric Hunting of New World Wild Sheep: Implications for the Academic Press. Honor of James Bennett Griffin, C. D. Cleland, ed. Pp. 293-318. New York: Study of Sheep Domestication. In Cultural Change and Continuity: Essays in

Wyatt, D. J.

1972 The Indian History of the Nicola Valley, British Columbia. Ph.D. Dissertation. Brown University, Providence.

Yesner, D. R.

1989 Moose Hunters of the Boreal Forest? A Re-examination of Subsistence Patterns in the Western Subarctic. Arctic 42:97-108.